

Temporal erosion of genetic diversity in Thailand's *Saccharum spontaneum* populations revealed by microsatellite markers

Anupong Wongtamee^{1,2,*}, Supansa Chinaworn³ and Tonapha Pusadee^{4,5}

¹Department of Agricultural Sciences, Faculty of Agriculture, Natural Resources and Environment, Naresuan University, Phitsanulok 65000, Thailand

²Center of Knowledge and Technology for Cane and Sugar, Faculty of agro-industry, Kasetsart University, 50 Ngamwongwan Road, Ladyao, Chatuchak, Bangkok 10900 Thailand

³Department of Plant Production Technology, Faculty of Agriculture and Natural Resources, Rajamangala University of Technology Tawan-Ok, Bangpra Campus, Chon buri 20110, Thailand

⁴Department of Plant and Soil Sciences, Faculty of Agriculture, Chiang Mai University, Chiang Mai 50200, Thailand

⁵Agrobiodiversity in Highland and Sustainable Utilization Research Group, Faculty of Agriculture, Chiang Mai University, Chiang Mai 50200, Thailand

*Corresponding author: anupongw@nu.ac.th

Abstract: Understanding temporal shifts in genetic diversity is critical for maintaining wild sugarcane germplasm. This study used 30 simple sequence repeat (SSR) loci to investigate genetic erosion in *Saccharum spontaneum* populations in lower northern Thailand over three collection periods (2017, 2019 and 2023). A total of 40 populations representing 15 sampling locations across four provinces in lower northern Thailand were analyzed. These populations were selected to encompass the major lowland habitats of *Saccharum spontaneum* (floodplains, riverbanks, canals, and swamps), providing an ecologically representative subset of Thai wild sugarcane populations for temporal genetic analysis. Polymorphism, gene diversity, and population structure were investigated in 40 populations across four provinces. The findings revealed a gradual decline in genetic diversity over time, consistent with demographic contraction and habitat fragmentation rather than directional selection, with polymorphic loci (PL) and Nei's gene diversity (h) dropping by 15-20%. When compared to 2017 and 2019, the 2023 populations had lower heterozygosity and narrower allelic richness. AMOVA analysis revealed that 50-52% of total variation occurred within populations, while 33-37% was partitioned among collection years and habitat-size classes ($\Phi_{PT} = 0.33-0.37$; $P < 0.001$). This indicates moderate genetic differentiation. Neighbor-joining clustering revealed two major temporal groups (2017-2019 vs. 2023), which are consistent with habitat fragmentation and limited gene flow. These findings provide the first temporal evidence of genetic erosion in Thai *S. spontaneum* germplasm, emphasizing the importance of targeted *in situ* and *ex situ* conservation, as well as incorporating wild allelic resources into sugarcane breeding programs under the BCG framework.

Keywords: *Saccharum spontaneum*, genetic erosion, SSR-based diversity, population structure, conservation in lower northern Thailand

Introduction

One of the most significant industrial crops in the world today is sugarcane (*Saccharum* spp. hybrids), which yields sugar, bioethanol, and renewable energy. Its limited genetic base, however, is mostly the result of a few interspecific crosses between the wild, stress-resistant *S. spontaneum* and the high-sucrose *S. officinarum* (Dillon et al., 2007; Zhang et al., 2018). While *S. officinarum* contributes sugar content, *S. spontaneum* provides important adaptive traits like drought, salinity, and disease tolerance, as well as tillering vigor and ratooning capacity (Meng et al., 2020). Thus, *S. spontaneum* may be a genetic reservoir for improving commercial sugarcane stress tolerance.

According to Ha et al. (1999) and Raboin et al. (2006), Thailand is known as a secondary center of diversity for *S. spontaneum*, with populations dispersed throughout a broad ecological gradient, from roadside wetlands to floodplains and riverbanks, and exhibiting exceptional adaptability. They provide valuable genetic resources for national and regional cane breeding programs. They are threatened by the continuous loss of natural habitats due to infrastructure development, land conversion, and agricultural expansion (Hasan et al., 2020). Long-term evolutionary potential is threatened by habitat fragmentation because it can lead to increased genetic drift, hinder gene flow, and ultimately decrease allelic richness within isolated populations (Aguilar et al., 2008; Frankham, 2015).

Previous research has used SSR or AFLP markers to characterize the genetic diversity of *S. spontaneum* in specific regions (Liu et al., 2016; Govindaraj et al., 2021), but these studies were largely cross-sectional, capturing spatial variation at a single time point rather than assessing temporal changes in diversity within the same populations. The rate and magnitude of genetic erosion in *S. spontaneum* under real-world disturbances remain unknown, particularly in Southeast Asia, where urbanization and agro-industrial expansion are most pronounced. Quantifying such temporal patterns is essential for detecting early warning signs of genetic depletion and developing effective conservation strategies before irreversible loss occurs.

To fill this gap, the current study uses 30 highly polymorphic microsatellite (SSR) markers to look at temporal changes in the genetic diversity and structure of *S. spontaneum* populations in lower northern Thailand over three collection years (2017, 2019, and 2023). Specifically, we wanted to (i) quantify changes in allelic richness and heterozygosity across time and population size classes, and (ii) assess how habitat loss and fragmentation affect within- and between-population genetic differentiation. We hypothesized that (1) allelic diversity in *S. spontaneum* has decreased over time as a result of ongoing habitat degradation, and (2) smaller or fragmented populations have greater genetic differentiation (higher *FST*) than larger, contiguous populations.

This study is the first to use spatial-temporal analyses in conjunction with molecular evidence to assess genetic erosion in Thailand's wild sugarcane gene pool over a period of years. The findings provide practical insights for conservation planning and pre-breeding utilization as part of Thailand's Bio-Circular-Green (BCG) economy initiative.

Results

SSR polymorphism and marker informativeness

Across 40 *S. spontaneum* populations genotyped with 30 microsatellite loci, 426 alleles were found (Table 2). The number of alleles per locus ranged from 14 (SMC703BS, SEGM6) to 20 (SMC278CS, mSSCIR3), with an overall mean of 16 loci⁻¹, confirming the high polymorphism typical of *Saccharum* polyploids. The average allele frequency was 0.83, and the polymorphic information content (PIC) values ranged from 0.32 to 0.63 (mean = 0.44), indicating that all loci were moderately to highly informative. Twelve markers (40%), notably SMC278CS, mSSCIR3, and CIR43, had PIC > 0.50 and were the most effective at detecting inter-population variation. Replicated genotyping yielded ≥98% concordance, demonstrating data reliability.

Table 1. Summary of 15 *S. spontaneum* populations surveyed in four provinces (Phitsanulok, Sukhothai, Phichit, and Nakhon Sawan) of lower northern Thailand across three collection years (2017, 2019, and 2023). The table presents habitat type and population size (m²) estimated from geospatial mapping.

No.	Population ID	Province	Habitat type	2017 (m ²)	2019 (m ²)	2023 (m ²)
1	PSL1	Phitsanulok	Deep swamp	253,280	198,362	101,540
2	PSL2	Phitsanulok	Roadside canal	1,065	872	672
3	PSL3	Phitsanulok	Swamp	2,852	2,014	1,462
4	PSL4	Phitsanulok	Floodplain	8,085	2,796	0 (lost to agriculture)
5	PSL5	Phitsanulok	Roadside canal	2,005	1,842	1,704
6	SKT1	Sukhothai	Canal	1,758	1,604	1,452
7	SKT2	Sukhothai	Floodplain	2,098	1,475	0 (lost to agriculture)
8	SKT3	Sukhothai	Roadside canal	894	714	598
9	PC1	Phichit	Deep swamp	128,792	0 (converted to reservoir)	0
10	PC2	Phichit	Floodplain	1,086	684	0 (lost to agriculture)
11	PC3	Phichit	Roadside canal	640	514	386
12	PC4	Phichit	Riverbank	768	702	568
13	PC5	Phichit	Roadside canal	646	580	488
14	NSW1	Nakhon Sawan	Deep swamp	375,160	301,753	265,410
15	NSW2	Nakhon Sawan	Riverbank	2,014	1,842	1,484

Population area was calculated from GPS coordinates using Google Earth Pro (v7.3). A value of "0" indicates complete habitat loss due to agricultural expansion or reservoir construction.

Table 2. Summary of polymorphism statistics for 30 SSR loci used to assess genetic diversity in *Saccharum spontaneum* populations. Markers with PIC ≥ 0.50 were considered highly informative for detecting inter-population variation.

No.	SSR marker	Expected fragment size (bp)	No. of alleles	Mean frequency	allele	PIC value
1	SMC119CG	95–192	19	0.87		0.58
2	SMC1604SA	98–145	16	0.81		0.42
3	SMC1751CL	123–175	17	0.84		0.47
4	SMC18SA	135–163	16	0.83		0.42
5	SMC22DUQ	124–174	18	0.89		0.53
6	SMC24DUQ	123–159	17	0.85		0.47
7	SMC278CS	111–225	20	0.91		0.63
8	SMC319CG	90–223	15	0.76		0.37
9	SMC31CUQ	127–215	16	0.83		0.42
10	SMC334BS	137–185	15	0.78		0.37
11	SMC336BS	137–222	16	0.89		0.42
12	SMC36BUQ	99–291	15	0.81		0.37
13	SMC486CG	187–301	15	0.72		0.37
14	SMC569CS	131–252	14	0.69		0.32
15	SMC597CS	109–247	15	0.83		0.37
16	SMC703BS	187–234	14	0.72		0.32
17	SMC7CUQ	99–253	16	0.67		0.42
18	SMC851MS	111–171	18	0.87		0.53
19	CIR3	159–355	18	0.91		0.53
20	CIR43	148–253	19	0.94		0.58
21	CIR66	101–155	16	0.81		0.42
22	CIR74	105–244	15	0.79		0.37
23	mSSCIR3	146–192	20	0.95		0.63
24	mSSCIR43	183–252	18	0.91		0.53
25	mSSCIR66	127–191	16	0.82		0.42
26	mSSCIR74	223–272	17	0.94		0.47
27	SEGM1	131–246	15	0.83		0.37
28	SEGM4	133–244	16	0.78		0.42
29	SEGM6	83–184	14	0.82		0.32
30	SEGM9	167–347	17	0.78		0.47
Mean $\pm$ SD	—	—	16 $\pm$ 2	0.83 $\pm$ 0.08		0.44 $\pm$ 0.09

PIC, polymorphic information content; bp, base pairs. Values were calculated from 30 SSR loci analyzed across 40 *S. spontaneum* populations from four provinces of lower northern Thailand.

Temporal and spatial patterns of genetic diversity

Overall, there was a high level of within-population diversity (PL = 0.601; $h = 0.565$; $H_s = 0.503$; $H_t = 0.838$), but there was only moderate differentiation between populations ($F_{ST} = 0.400$). However, diversity indices declined steadily from 2017 to 2023 (Figure 2). In 2017, populations had the highest polymorphism (PL = 0.714) and gene diversity ($h = 0.597$), but both parameters decreased slightly in 2019 (PL = 0.633; $h = 0.569$). By 2023, PL and h had dropped to 0.457 and 0.529, respectively, representing a 12–18% reduction from 2017 levels.

The temporal erosion was accompanied by a modest rise in population differentiation ($F_{ST} = 0.388 \rightarrow 0.419$), indicating increased genetic isolation among remaining stands. The loss of allelic richness coincided with a documented reduction in population size (Table 1), implying that demographic contraction was the most likely driver.

Spatially, provincial comparisons revealed the greatest genetic diversity in Phitsanulok ($h = 0.658$ in 2017) and Nakhon Sawan ($h = 0.587$ in 2023), reflecting their larger, floodplain-connected habitats. Phichit populations, many of which experienced dredging or conversion to agriculture, showed significant declines ($h \approx 0.51$ in 2023). F-tests revealed significant temporal variation in PL in Phitsanulok ($P < 0.05$), but not in other provinces, indicating site-specific erosion.

Influence of population size

Large populations ($> 10,000 \text{ m}^2$) showed the highest allelic richness and heterozygosity across all years (Table 3). In 2023, small populations ($< 1,000 \text{ m}^2$) showed slightly higher average gene diversity ($H_s = 0.510$) compared to medium populations ($H_s = 0.502$), indicating that local regeneration or short-distance gene flow may temporarily buffer drift effects.

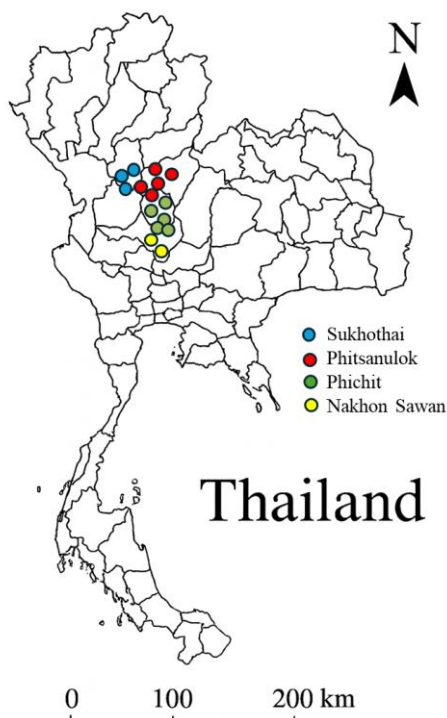


Figure 1. Map of Thailand highlighting the study area in lower northern Thailand, showing the locations of 15 *Saccharum spontaneum* populations (PSL1–PSL5 in Phitsanulok, SKT1–SKT3 in Sukhothai, PC1–PC5 in Phichit, and NSW1–NSW2 in Nakhon Sawan) surveyed during 2017–2023. The mapped region represents a core lowland distribution area of wild *Saccharum* rather than the entire national distribution. Colored circles represent provinces: red for Phitsanulok, blue for Sukhothai, green for Phichit, and yellow for Nakhon Sawan.

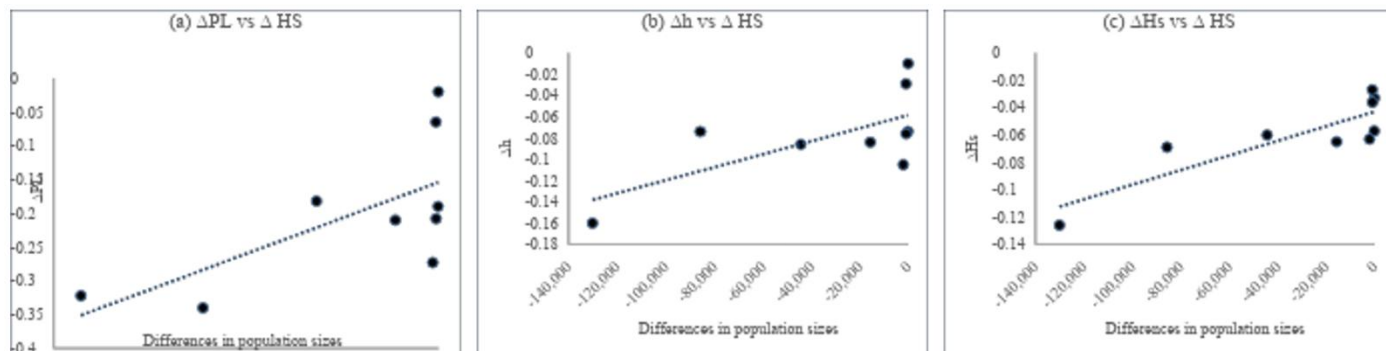


Figure 2. Relationships between changes in habitat size (ΔHS) and genetic diversity indices of *S. spontaneum* populations across three collection periods (2017–2019, 2019–2023, and 2017–2023): (a) ΔPL vs ΔHS , (b) Δh vs ΔHS , and (c) ΔH_s vs ΔHS . Each point represents one population size class. Dashed regression lines and corresponding R^2 values indicate that decreasing habitat area is associated with reduced genetic diversity.

Despite this anomaly, ANOVA revealed significant differences in PL and H_s among size classes ($P < 0.05$), confirming population size as an important but not exclusive predictor of genetic variability.

Temporal shifts in genetic structure

Neighbor-Joining clustering using Nei's (1983) genetic distance separated all accessions into two major clusters (Figure 3a), with medium 2023 populations forming a secondary sub-lineage (Figure 3b). Cluster I contained 26 populations sampled in 2017 and 2019, primarily from Phitsanulok, Sukhothai, and Phichit, whereas Cluster II contained 14 populations sampled in 2023, including several small or fragmented sites.

This grouping represents a temporal restructuring of allele composition, with 2023 populations forming a distinct sub-lineage with higher internal homogeneity. The pattern was partially replicated when populations were classified by size, with medium 2023 populations clustering together, implying that medium-sized habitats may be more vulnerable to directional drift.

Partitioning of molecular variance (AMOVA)

AMOVA confirmed that within-population variance predominated, accounting for 50.0% and 52.0% of total variance in temporal and size-based groups, respectively (Table 4). Among-population components accounted for 33-37%, which is consistent with the moderate structuring found in outcrossing grasses. Although the proportion of within-population

Table 3. Genetic diversity indices of *S. spontaneum* populations across collection years, provinces, and population-size classes in lower northern Thailand (2017–2023). A temporal decline in PL and *h* and moderate differentiation ($F_{ST} \approx 0.40$) indicate progressive genetic erosion linked to habitat fragmentation.

Year / Category	No. of populations	PL	<i>h</i>	<i>H_s</i>	<i>H_t</i>	F_{ST}
Overall (2017–2023)	40	0.601 ± 0.04	0.565 ± 0.03	0.503 ± 0.02	0.838 ± 0.02	0.4
Year of collection						
2017	15	0.714 ± 0.03	0.597 ± 0.02	0.531 ± 0.02	0.867 ± 0.01	0.388
2019	14	0.633 ± 0.04	0.569 ± 0.03	0.507 ± 0.02	0.837 ± 0.02	0.395
2023	11	0.457 ± 0.06	0.529 ± 0.04	0.471 ± 0.03	0.810 ± 0.03	0.419
Province (2017–2023)						
Phitsanulok	5	0.606 ± 0.07	0.597 ± 0.05	0.531 ± 0.03	0.861 ± 0.03	0.384
Sukhothai	3	0.525 ± 0.06	0.530 ± 0.04	0.466 ± 0.03	0.773 ± 0.02	0.39
Phichit	5	0.567 ± 0.05	0.563 ± 0.04	0.506 ± 0.03	0.881 ± 0.03	0.427
Nakhon Sawan	2	0.695 ± 0.03	0.617 ± 0.03	0.541 ± 0.02	0.878 ± 0.02	0.385
Population size class (m ²)						
< 1 000	4	0.603 ± 0.04	0.564 ± 0.03	0.510 ± 0.03	0.827 ± 0.03	0.383
1 001–10 000	7	0.546 ± 0.05	0.556 ± 0.03	0.502 ± 0.02	0.860 ± 0.02	0.419
> 10 000	3	0.619 ± 0.04	0.606 ± 0.03	0.526 ± 0.02	0.896 ± 0.02	0.392

PL = proportion of polymorphic loci; *h* = Nei's (1973) gene diversity; *H_s* = average within-population diversity; *H_t* = total gene diversity; F_{ST} = genetic differentiation among populations. Values are means ± standard deviations computed from 30 SSR loci across 40 populations.

Table 4. Analysis of molecular variance (AMOVA) of *S. spontaneum* populations across collection years (2017–2023) and habitat-size classes. Within-population variance (~50%) indicates strong individual heterogeneity, while moderate Φ_{PT} values (0.33–0.37; $P < 0.001$) reveal temporal and spatial genetic structuring due to restricted gene flow.

Source of variation	df	SS	MS	Estimated variance	% of total variation
(a) Temporal grouping (2017, 2019, 2023)					
Among collection years	2	350.1	175.0	0.37	37.0
Among populations within years	72	130.4	1.81	0.13	13.0
Within populations	75	470.2	6.27	0.50	50.0
Total	149	950.7	–	1.00	100.0
Φ_{PT} (among years vs. total)					0.37***
(b) Habitat-size grouping (<1,000; 1,001–10,000; >10,000 m ²)					
Among size classes	2	310.5	155.3	0.33	33.0
Among populations within size classes	72	144.9	2.01	0.15	15.0
Within populations	75	480.8	6.41	0.52	52.0
Total	149	936.2	–	1.00	100.0
Φ_{PT} (among size classes vs. total)					0.33***

df = degrees of freedom; SS = sum of squares; MS = mean squares; Φ_{PT} = fixation index measuring population differentiation. All tests were based on 999 random permutations using GenAlEx v6.1. Significance levels: $P < 0.001$ (***).

variance remained high, its relative contribution decreased over time, confirming the genetic differentiation trend predicted by the F_{ST} . Thailand's *S. spontaneum* populations lost genetic diversity over six years, most likely due to habitat fragmentation and population decline.

Discussion

High intrinsic diversity and evolutionary mechanisms

S. spontaneum's high allelic richness and heterozygosity make it a genetically dynamic grassland and riparian species in Thailand. This diversity stems from the polyploid genome architecture and predominantly out-crossing reproductive system, both of which improve allelic buffering and preserve multi-locus polymorphism under selection pressure (Liu et al.,

2016; Medeiros et al., 2020). Polyploid buffering lets duplicate loci accumulate mutations without fitness penalties, preserving long-term adaptive potential. Even though floodplain populations were isolated, historical gene flow likely maintained high within-population variance (50-52%). These findings support the idea that *S. spontaneum* is an irreplaceable genetic reservoir for stress-tolerance traits, such as drought and salinity resistance, that can be introduced into commercial sugarcane lines (Tai and Miller, 2002; Liu et al., 2022). Despite the genomic complexity of *Saccharum*, the use of a moderate number of highly informative SSR loci has been widely shown to capture population-level genetic diversity effectively in polyploid species. The strong concordance between

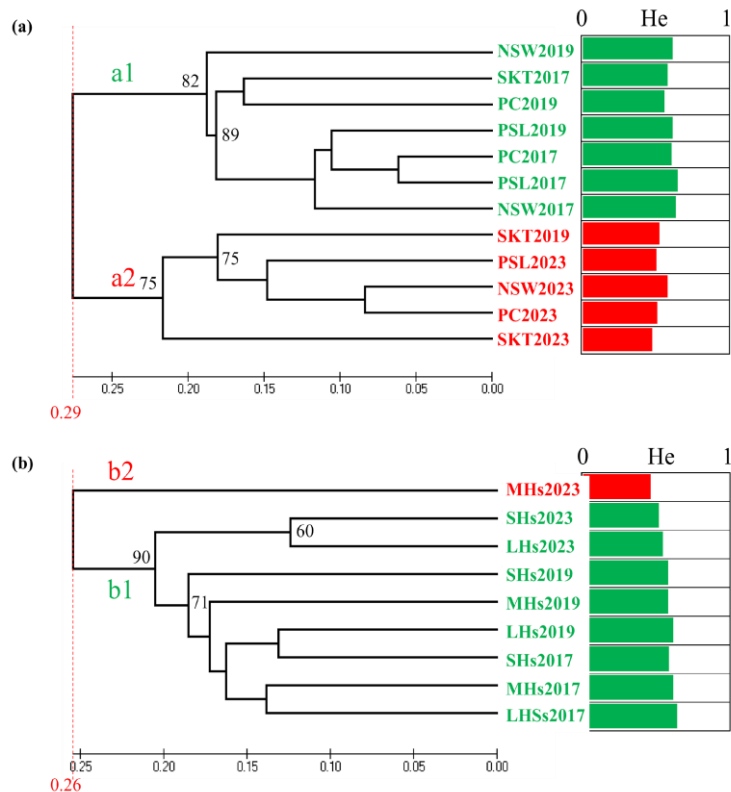


Figure 3. Neighbor-Joining (NJ) cluster analysis of 40 *S. spontaneum* populations from four provinces in lower northern Thailand based on Nei's (1983) genetic distance. (a) Populations grouped by collection years (2017, 2019, 2023). (b) Populations grouped by habitat size classes: small (SHs; <1,000 m²), medium (MHs; 1,001–10,000 m²), and large (LHs; >10,000 m²). Bootstrap values (1,000 replicates) are shown at all nodes, and data bars represent expected heterozygosity (He), highlighting temporal and size-related genetic differentiation.

temporal trends in PL, h, FST, AMOVA partitioning, and clustering patterns observed in this study indicates that the selected marker set reliably reflects underlying evolutionary processes rather than sampling artefacts.

Temporal genetic erosion and its underlying drivers

The gradual decline in genetic diversity observed between 2017 and 2023 is unlikely to be primarily driven by directional natural selection. The SSR loci employed in this study are assumed to be selectively neutral; therefore, changes in diversity metrics are more plausibly attributed to demographic and stochastic processes. Progressive habitat loss, population size reduction, and fragmentation reduce effective population size and restrict pollen- and seed-mediated gene flow, thereby intensifying genetic drift and accelerating allelic loss.

The concurrent decrease in polymorphic loci (PL) and Nei's gene diversity (h), together with increasing FST values and greater among-group variance in AMOVA, supports a drift-dominated erosion model rather than selection-driven differentiation. While local environmental pressures may impose selection on adaptive traits, such effects are unlikely to be detected by neutral SSR markers. Thus, the observed temporal genetic erosion primarily reflects landscape-mediated demographic processes acting on remnant populations.

This study shows molecular genetic diversity erosion between 2017 and 2023 despite its robustness. Polymorphic locus reductions of 12-18%, expected heterozygosity, and higher FST values indicate genetic drift and restricted gene flow in fragmented habitats. AMOVA supports Wright's (1978) interpretation by revealing a 33-37% variation between collection years and habitat-size classes ($\Phi_{PT} = 0.33-0.37$; $P < 0.001$). This suggests moderate differentiation among outcrossing taxa. The decline corresponds to regional land-use transitions documented in Phichit and Phitsanulok provinces at the same time, such as canal dredging, paddy expansion, and road construction. These disturbances reduce population size and continuity, hastening drift and encouraging allele fixation in small remnants (Hedrick et al., 2014; Miles et al., 2019). Surprisingly, a slight increase in diversity among some small 2023 populations suggests transient recruitment from seed banks or short-distance dispersal, which partially offsets drift. Following flooding events, *Erianthus arundinaceus* and *Miscanthus sinensis* have shown similar short-term rebounds (Wang et al., 2008). However, if fragmentation persists, stochastic regeneration cannot counteract long-term erosion, emphasizing the importance of continuous monitoring.

Population size, drift-gene-flow equilibrium, and spatial heterogeneity

The complex relationship between population size and diversity observed here lends support to the drift-gene-flow equilibrium model. Large floodplain populations maintained high heterozygosity due to extensive cross-pollination, whereas medium-sized patches experienced greater allele loss, most likely due to being large enough for drift to act but too isolated for effective gene exchange. They support our hypothesis that smaller or fragmented populations have stronger genetic differentiation due to restricted gene flow. Similar nonlinear patterns have been observed in *Phragmites australis* and *Oryza rufipogon* (Leimu et al., 2006; Wongtamee et al., 2017). As a result, conservation strategies must acknowledge that "moderate isolation" can be more genetically dangerous than either extensive connectivity or very small but frequently recolonized patches.

Spatial contrasts also highlight the importance of landscape connectivity: Nakhon Sawan's vast wetlands maintained high diversity even in 2023, whereas Phichit's channelized systems experienced the steepest decline. Maintaining hydrological linkages and riparian corridors will help gene flow between populations.

Comparisons with global *Saccharum* diversity studies

Compared to earlier single-time surveys in India (Govindaraj et al., 2021) and China (Liu et al., 2022), Thai *S. spontaneum* populations still exhibit comparable or higher allelic richness ($A \approx 16 \text{ locus}^{-1}$), but with a greater temporal fluctuation, suggesting recent anthropogenic impact rather than historical bottleneck. The current study found moderate differentiation ($\Phi_{PT} \approx 0.40$), similar to *Erianthus* spp. in Southeast Asia (Ali et al., 2019), indicating that wild *Saccharum* taxa in tropical Asia exhibit similar levels of within-population variation but differ in temporal stability. The current study fills a regional gap by demonstrating that genetic erosion can occur within six years, or one to two clonal generations, emphasizing polyploid perennials' sensitivity to habitat disturbance.

Conservation and breeding implications

The documented decline is an early indicator of germplasm vulnerability in Thailand's wild sugarcane gene pool. Conservation measures should prioritize the following: (1) *in situ* protection of the most diverse floodplain and canal-bank habitats (e.g., Phitsanulok, Nakhon Sawan). (2) *Ex situ* preservation using living collections and cryopreserved germplasm banks to capture allelic diversity over multiple years. (3) Create connecting riparian corridors or buffer strips for pollen and seed flow. (4) Tracking *He* and *FST* diversity metrics with molecular monitoring in the national sugarcane breeding network. Thailand's Bio-Circular-Green (BCG) economic framework promotes sustainable use of native biodiversity, which supports these actions. *S. spontaneum* alleles in pre-breeding pipelines can improve commercial cultivars' drought, salinity, and pest resistance, which are becoming increasingly important due to climate variability. Due to moderate differentiation ($\Phi_{PT} \approx 0.33\text{-}0.37$), 15-20% loss of polymorphism, and 12-18% decline in gene diversity, fragmented habitats rapidly lose allelic richness, highlighting the need for coordinated conservation interventions.

Future perspectives

To identify environmental variables that predict biodiversity loss, future research should combine molecular monitoring with landscape genomics and remote sensing analyses. High-throughput SNP genotyping and genome-wide association mapping may reveal candidate genes involved in hydrological stress adaptation. Extending temporal sampling beyond 2023 will reveal whether observed declines are a one-time occurrence or a long-term trend toward genetic impoverishment. Evidence-based conservation planning for *Saccharum* wild relatives in mainland Southeast Asia will require genetic, ecological, and socioeconomic data.

Material and methods

Study area and population sampling

Field surveys of *S. spontaneum* populations were conducted in four provinces in lower northern Thailand: Phitsanulok (PSL), Sukhothai (SKT), Phichit (PC), and Nakhon Sawan (NSW), where *S. spontaneum* is commonly found in floodplains, riverbanks, and roadside wetlands (Figure 1). The tropical monsoon climate and seasonal flooding characterize the Nan-Yom-Ping River basin's lowland alluvial ecosystems. Although geographically focused, the study area represents a major lowland ecological zone where wild *Saccharum spontaneum* is widespread in Thailand and frequently exposed to similar land-use pressures.

Sampling campaigns were carried out at three time points (2017, 2019, and 2023) to assess temporal changes in population diversity. In total, 40 populations derived from 15 sampling locations were included in this study. Sampling sites were distributed across four provinces in lower northern Thailand (Phitsanulok, Sukhothai, Phichit, and Nakhon Sawan), covering the major ecological settings in which wild *Saccharum spontaneum* naturally occurs. Site selection was guided by explicit criteria to ensure ecological and genetic representativeness, including (i) geographic separation of populations ($>15 \text{ km}$), (ii) minimum habitat size ($>100 \text{ m}^2$ at initial sampling), (iii) inclusion of contrasting habitat types (floodplains, riverbanks, canals, and swamps), and (iv) persistence of populations across multiple collection years to allow temporal comparisons.

Although the sampling was geographically focused on lower northern Thailand, this region constitutes a core distribution area of wild *Saccharum* in the country. Therefore, the selected populations provide a representative model for assessing temporal genetic erosion in Thai lowland *Saccharum spontaneum* populations under increasing anthropogenic pressure.

To reduce clonal redundancy, 50 plants were randomly sampled at ≥ 2 m spacing from each site. Leaf tissues were collected, desiccated in silica gel, and stored at -20 degrees Celsius until DNA extraction. Table 1 shows the sampling coordinates, habitat type, and population size for each collection.

DNA extraction and SSR genotyping

Genomic DNA was extracted from silica-dried leaves using a modified CTAB protocol (Pan, 2006), which was optimized for polysaccharide-rich grass tissues. DNA quality and quantity were determined spectrophotometrically (NanoDrop 2000, Thermo Scientific) and confirmed with 1% agarose gel electrophoresis.

Thirty previously validated simple sequence repeat (SSR) primer pairs for *Saccharum* (Table 2) were chosen due to their high polymorphism and genome coverage (Ali et al., 2019; Xiong et al., 2022). Although sugarcane possesses a highly complex polyploid genome, microsatellite markers remain appropriate for population-level analyses when the objective is to assess neutral genetic variation rather than full genome-wide coverage. The 30 SSR loci used in this study were highly polymorphic, multi-allelic, and distributed across different linkage groups, providing sufficient resolution to detect temporal and spatial genetic differentiation among populations.

PCR reactions (20 μ L) used 50 ng genomic DNA, 1 $\times$ buffer, 2 mM $MgCl_2$, 200 μ M dNTPs, 0.5 μ M primers, and 1 U Taq polymerase. The amplification program included 95 $^{\circ}C$ for 5 minutes, 35 cycles of 95 $^{\circ}C$ for 45 seconds, 55-60 $^{\circ}C$ for 45 seconds (primer-specific), 72 $^{\circ}C$ for 1 minute, and a 10-minute extension at 72 $^{\circ}C$.

Ethidium bromide-stained 10% polyacrylamide gels were used to examine amplified products under UV light. Each allele was measured using a 50 bp DNA ladder (GeneRuler™, Thermo Scientific) and scored in base pairs with GelAnalyzer v19.1. We achieved $\geq 98\%$ accuracy by re-amplifying and re-scoring 10% of samples to ensure consistency.

Genetic diversity analysis

Allele data were converted into a binary matrix (presence/absence) and analyzed using FSTAT v2.9.3 (Goudet, 2001) and GenAlEx v6.1 (Peakall and Smouse, 2006). The following parameters were computed; proportion of polymorphic loci (PL), number of alleles per locus (A), Nei's gene diversity (h) (Nei, 1973), average within-population diversity (Hs) and total diversity (Ht), and fixation index (Fst) indicating population differentiation.

Temporal comparisons were performed for 2017–2019, 2019–2023, and 2017–2023 periods. Populations were also grouped by habitat size (< 1,000 m², 1,001–10,000 m², > 10,000 m²) to test for size-related effects on diversity. Significance of parameter differences among years and population categories was assessed using one-way ANOVA and F-tests at $P < 0.05$.

Genetic structure and differentiation

Population relationships were assessed using Nei's (1983) genetic distance and Neighbor-Joining (NJ) clustering implemented in MEGA v11.0. Bootstrapping (1,000 replicates) was used to test branch stability. The proportion of molecular variance within and among populations was quantified through Analysis of Molecular Variance (AMOVA) in GenAlEx, based on 999 permutations.

To examine whether genetic differentiation corresponded to sampling year or population size, hierarchical AMOVA models were constructed for (i) temporal grouping (2017, 2019, 2023) and (ii) habitat-size grouping (small, medium, large).

Conclusion

This study provides the first temporal assessment of genetic diversity in *S. spontaneum* populations from lower northern Thailand. A consistent decline in polymorphic loci (PL) and Nei's gene diversity (h) from 2017 to 2023, together with AMOVA results showing ≈ 50 –52% within-population variance and Φ_{PT} values of 0.33–0.37 ($P < 0.001$), indicates progressive genetic erosion driven by habitat fragmentation and isolation. The two major clusters observed (2017–2019 vs. 2023) further confirm temporal restructuring of the gene pool. The results confirm our initial hypotheses that allelic diversity declined over time and that fragmentation intensified population differentiation.

These findings suggest urgent *in situ* and *ex situ* conservation of remnant *S. spontaneum* populations as allelic diversity reservoirs for BCG and climate-resilient agriculture sugarcane improvement.

Statement of contributions

Anupong Wongtamee conceived and designed the experiment, conducted field sampling, performed molecular and statistical analyses, and drafted the manuscript. Supansa Chinaworn contributed to statistical validation, data interpretation, and manuscript editing. Tonapha Pusadee supervised data visualization, critically revised the text, and approved the final version. All authors read and approved of the final manuscript.

Acknowledgements

This research was supported by Naresuan University (NU) and the National Science, Research and Innovation Fund (NSRF) under Grant No. R2566B078. The authors gratefully acknowledge the Faculty of Agriculture, Natural Resources and Environment, Naresuan University, and the Faculty of Agriculture, Chiang Mai University, for providing laboratory instruments, technical facilities, and field logistics. Special thanks are extended to the provincial agricultural officers and

community partners in Phitsanulok, Sukhothai, Phichit, and Nakhon Sawan for their invaluable assistance in sample collection and habitat documentation.

Conflict of interest

The authors declare that they have no conflict of interest.

References

- Aguilar R, Quesada M, Ashworth L, Herrerias-Diego Y, Lobo J (2008) Genetic consequences of habitat fragmentation in plant populations: susceptible signals in plant traits and methodological approaches. *Mol Ecol* 17(24):5177–5188. <https://doi.org/10.1111/j.1365-294X.2008.03971.x>
- Ali A, Pan YB, Wang QN, Wang JD, Chen JL, Gao SJ (2019) Genetic diversity and population structure analysis of *Saccharum* and *Erianthus* genera using microsatellite (SSR) markers. *Sci Rep* 9(1):395. <https://doi.org/10.1038/s41598-018-37233-2>
- Dillon SL, Shapter FM, Robert HJ, Cordeiro G, Izquierdo L, Lee SL (2007) Domestication to crop improvement: genetic resources for *Sorghum* and *Saccharum* (Andropogoneae). *Ann Bot* 5:975–989. <https://doi.org/10.1093/aob/mcm192>
- Frankham R (2015) Genetic rescue of small inbred populations: meta-analysis reveals large and consistent benefits of gene flow. *Mol Ecol* 24:2610–2618. <https://doi.org/10.1111/mec.13139>
- Goudet J (2001) FSTAT, a program to estimate and test gene diversities and fixation indices (version 2.9.3). [Internet]. Available from: <http://www2.unil.ch/popgen/softwares/fstat.htm>
- Govindaraj P, Gowri R, Mohanraj K, Amalraj VA (2021) SSR marker based molecular genetic diversity analysis among *Saccharum spontaneum* (L.) collected from north western region of India. *Sugar Tech* 23:730–740. <https://doi.org/10.1007/s12355-021-00956-w>
- Ha S, Moore PH, Heinz D, Kato S, Ohmido N, Fukui K (1999) Quantitative chromosome map of the polyploid *Saccharum spontaneum* by multicolor fluorescence in situ hybridization and imaging methods. *Plant Mol Biol* 39(6):1165–1173. <https://doi.org/10.1023/A:1006152103335>
- Hasan SS, Zhen L, Miah MG, Ahamed T, Samie A (2020) Impact of land use change on ecosystem services: a review. *Environ Dev* 34:100527. <https://doi.org/10.1016/j.envdev.2020.100527>
- Hedrick P, Peterson R, Vucetich L, Adams J, Vucetich J (2014) Genetic rescue in Isle Royale wolves: genetic analysis and the collapse of the population. *Conserv Genet* 15:1111–1121. <https://doi.org/10.1007/s10592-014-0604-1>
- Leimu R, Mutikainen P, Koricheva J, Fischer M (2006) How general are positive relationships between plant population size, fitness, and genetic variation? *J Ecol* 94:942–952. <https://doi.org/10.1111/j.1365-2745.2006.01150.x>
- Liu L, Wang H, Li Y, Chen S, Wu M, Dou M, Qi Y, Fang J, Zhang J (2022) Genome-wide development of interspecific microsatellite markers for *Saccharum officinarum* and *Saccharum spontaneum*. *J Integr Agric* 21(11):3230–3244. <https://doi.org/10.1016/j.jia.2022.08.129>
- Liu XL, Li XJ, Xu CH, Lin XQ, Deng ZH (2016) Genetic diversity of populations of *Saccharum spontaneum* with different ploidy levels using SSR molecular markers. *Sugar Tech* 18(4):365–372. <https://doi.org/10.1007/s12355-015-0320-x>
- Medeiros C, Balsalobre TWA, Carneiro MS (2020) Molecular diversity and genetic structure of *Saccharum* complex accessions. *PLoS One* 15(5):e0233211. <https://doi.org/10.1371/journal.pone.0233211>
- Meng Z, Han J, Lin Y, Zhao Y, Lin Q, Ma X, Wang J, Zhang M, Zhang L, Yang Q, Wang K (2020) Characterization of a *Saccharum spontaneum* L. with a basic chromosome number of $x = 10$ provides new insights into genome evolution in the genus *Saccharum*. *Theor Appl Genet* 133:187–199. <https://doi.org/10.1007/s00122-019-03448-z>
- Miles LS, Rivkin LR, Johnson MTJ, Munshi-South J, Verrelli BC (2019) Gene flow and genetic drift in urban environments. *Mol Ecol* 28(18):4138–4151. <https://doi.org/10.1111/mec.15221>
- Nei M (1973) The theory and estimation of genetic distance. In: Morton NE (ed) *Genetic structure of populations*. University of Hawaii Press, Honolulu, pp 45–54. Available from: <https://www.scirp.org/reference/referencespapers?referenceid=32519>
- Nei M (1983) Genetic polymorphism and the role of mutation in evolution. In: Nei M, Koehn RK (eds) *Evolution of genes and proteins*. Sinauer Associates, Sunderland, pp 165–190. <https://www.scirp.org/reference/referencespapers?referenceid=59366>
- Pan YB (2006) Highly polymorphic microsatellite DNA markers for sugarcane germplasm evaluation and variety identity testing. *Sugar Tech* 8:246–256. <https://doi.org/10.1007/BF02943564>
- Peakall R, Smouse PE (2006) GENALEX 6: genetic analysis in Excel. Population genetic software for teaching and research. *Mol Ecol Notes* 6(1):288–295. <https://doi.org/10.1111/j.1471-8286.2005.01155.x>
- Raboin LM, Oliveira KM, Lecunff L, Telismart H, Roques D, Butterfield M, Hoarau JY, D'Hont A (2006) Genetic mapping in sugarcane, a high polyploid, using bi-parental progeny: identification of a gene controlling stalk colour and a new rust resistance gene. *Theor Appl Genet* 112:1382–1391. <https://doi.org/10.1007/s00122-006-0240-3>
- Tai PYP, Miller JD (2002) Germplasm diversity among four sugarcane species for sugar composition. *Crop Sci* 42(3):958–964. <https://doi.org/10.2135/cropsci2002.9580>
- Wang LP, Jackson PA, Lu X (2008) Evaluation of sugarcane $\times$ *Saccharum spontaneum* progeny for biomass composition and yield components. *Crop Sci* 48:951–961. <https://doi.org/10.2135/cropsci2007.10.0555>

- Wongtamee A, Maneechote C, Pusadee T, Rerkasem B, Jamjod S (2017) The dynamics of spatial and temporal population genetic structure of weedy rice (*Oryza sativa* f. *spontanea* Baker). *Genet Resour Crop Evol* 64:1073–1087. <https://doi.org/10.1007/s10722-015-0330-7>
- Xiong H, Chen Y, Gao S, Pan Y, Shi A (2022) Population structure and genetic diversity analysis in sugarcane (*Saccharum* spp. hybrids) and six related *Saccharum* species. *Agronomy* 12(2):412. <https://doi.org/10.3390/agronomy12020412>
- Zhang J, Zhang X, Tang H, Zhang Q, Hua X, Ma X, et al. (2018) Allele-defined genome of the autopolyploid sugarcane *Saccharum spontaneum* L. *Nat Genet* 50(11):1565–1573. <https://doi.org/10.1038/s41588-018-0237-2>
- Wright S (1978) *Evolution and the genetics of populations, Vol. 4: Variability within and among natural populations*. University of Chicago Press, Chicago, IL.